

Research review

Evolutionary ecology of resprouting and seeding in fire-prone ecosystems

Author for correspondence:

Juli G. Pausas

Tel: +34 963 424124

Email: juli.g.pausas@uv.es

Received: 2 January 2014

Accepted: 26 May 2014

Juli G. Pausas¹ and Jon E. Keeley^{2,3}

¹CIDE-CSIC, Ctra. Naquera Km 4.5 (IVIA), 46113 Montcada, Valencia, Spain; ²US Geological Survey, Sequoia-Kings Canyon Field Station, Three Rivers, CA 93271, USA; ³Department of Ecology and Evolutionary Biology, University of California, Los Angeles, CA 90095, USA

New Phytologist (2014) **204**: 55–65
doi: 10.1111/nph.12921

Key words: fire ecology, iteroparity, monopyric, polypyric, resprouting, seeding, semelparity, trait evolution.

Summary

There are two broad mechanisms by which plant populations persist under recurrent disturbances: resprouting from surviving tissues, and seedling recruitment. Species can have one of these mechanisms or both. However, a coherent framework explaining the differential evolutionary pressures driving these regeneration mechanisms is lacking. We propose a bottom-up approach in addressing this question that considers the relative survivorship of adults and juveniles in an evolutionary context, based on two assumptions. First, resprouting and seeding can be interpreted by analogy with annual versus perennial life histories; that is, if we consider disturbance cycles to be analogous to annual cycles, then resprouting species are analogous to the perennial life history with iteroparous reproduction, and obligate seeding species that survive disturbances solely through seed banks are analogous to the annual life history with semelparous reproduction. Secondly, changes in the selective regimes differentially modify the survival rates of adults and juveniles and thus the relative costs and benefits of resprouting versus seeding. Our approach provides a framework for understanding temporal and spatial variation in resprouting and seeding under crown-fire regimes. It accounts for patterns of coexistence and environmental changes that contribute to the evolution of seeding from resprouting ancestors.

Introduction

Recurrent disturbances underpin the dynamics of many ecosystems world-wide and exert a strong evolutionary pressure on plants. Many species have consequently acquired traits and mechanisms that confer fitness benefits in repeatedly disturbed environments. There are two broad mechanisms by which plant populations persist in the face of recurrent disturbances: resprouting from surviving parental tissues and seedling recruitment (Bond & Midgley, 2001; Pausas *et al.*, 2004). Despite attempts to explain the relative roles of these two regeneration mechanisms (Bellingham & Sparrow, 2000; Bond & Midgley, 2001, 2003; Pausas, 2001; Klimesová & Klimes, 2003, 2007; Pausas *et al.*, 2004; Vesik & Westoby, 2004; Knox & Clarke, 2005; Pausas & Bradstock, 2007; Lamont *et al.*, 2011; Clarke *et al.*, 2013), a coherent framework explaining the differential evolutionary pressures driving them is still lacking.

We propose a bottom-up approach to address the relative roles of resprouting and seeding that considers relative survivorship of

adults and juveniles in an evolutionary context. We focus on ecosystems where fire is the primary disturbance and frequent enough to act as a strong selective pressure (fire-prone ecosystems), and specifically on ecosystems where fire typically kills all aboveground plant parts (crown-fire ecosystems; Box 1). Our approach is based on two assumptions. The first assumption is that resprouting and seeding can be interpreted by analogy with annual versus perennial life histories. If we consider fire cycles to be analogous to annual cycles, then by analogy obligate seeding species that survive fires solely through seed banks are like annual species with semelparous life histories (Keeley, 1986; Bond & Van Wilgen, 1996; monopyric species; Box 1; Table 1). By contrast, resprouters persist through multiple fire cycles and are analogous to perennial species (polypyric species; Box 1) with typically iteroparous reproduction across multiple cycles (Table 1). Despite occasional reproductive events without fire in some obligate seeding species (Nathan *et al.*, 1999), the bulk of the effective reproduction in these species occurs after a fire when the plant dies, and thus they can effectively be considered semelparous with a single reproductive

Box 1 Basic concepts

Postfire regeneration traits

- **Postfire resprouting:** the ability to generate new shoots from dormant buds after stems have been fully scorched by fire. This term is preferable to *sprouting*, which refers to initiation of new shoots throughout the life cycle of a plant. Species are typically classified as resprouters or nonresprouters depending on this ability. There are different types of resprouting depending on the location of the buds (epicormic, lignotuber, rhizome, roots, etc.)
- **Postfire seeding:** the ability to generate a fire-resistant seed bank with seeds that germinate profusely after fires (fire-cued germination). Typically, such species restrict recruitment to a single pulse after a fire. Seeds may be stored in the soil or in the canopy (seed bank; Box 3). Species are typically classified as seeders or nonseeders (or fire-dependent/fire-independent recruiters) depending on this ability. There are different types of postfire seeding (Box 3).

Postfire strategies

- **Obligate resprouters:** plants that rely on resprouting to regenerate after fire (resprouters without postfire seeding ability). These plants do not germinate after fire because they lack a fire-resistant seed bank. Note that obligate resprouters might reproduce by seeds during the fire-free interval, but the terminology of seeders and resprouters refers to the postfire conditions.
- **Facultative seeders:** plants that have both mechanisms for regenerating after fire, that is, they are able to resprout and to germinate after fire.
- **Obligate seeders:** plants that do not resprout and rely on seeding to regenerate their population after fire (nonresprouters with postfire seeding ability). Because they tend to recruit massively once in their lifespan (after fire) and fire kills the adults and typically exhausts their seed bank, they can be considered semelparous species with nonoverlapping generations and a monopyric life cycle (Table 1). Note that the term 'seeders' refers strictly to postfire conditions, and cannot be attributed to plants that regenerate by seeds in other conditions.
- **Postfire colonizers:** plants that lack a mechanism for local postfire persistence, but they recruit after fire by seeds dispersed from unburned patches or from populations outside the fire perimeter (metapopulation dynamics).

Life cycle in relation to fire

We propose the following terminology to define the life cycle of an organism living in a fire prone ecosystem:

- **Monopyric:** species that perform all their life cycle within a fire cycle. In plants, examples are annual and biennial species, postfire obligate seeders and some bamboos.
- **Polypyric:** species that perform all their life cycle through multiple fire cycles. In plants, examples are those with postfire resprouting capacity as well as trees with other survival strategies such as very thick bark.

Basic fire regimes

- **Surface fires:** fires that spread in the herbaceous or litter layer, such as the understory of some forests and in savannas and grasslands. These fires are usually of relatively low intensity and high frequency.
- **Crown-fires:** fires in woody-dominated ecosystems that affect all vegetation including crowns. They are typically of high intensity. Examples are fires in some Mediterranean-type forest and shrublands and in closed-cone pine forests.

form even-aged populations while resprouting species form multiple-cohort populations) and in the genetic variability and the evolutionary potential (obligate seeding species have shorter generation time and faster population turnover). Our second assumption is that changes in the selective regimes differentially modify the survival rates of adults and juveniles and thus the relative costs and benefits of resprouting versus seeding. These changes would drive relaxed selection of one trait (i.e. resprouting), increasing the probability of its loss (Lahti *et al.*, 2009), and the intensification of the selection of another trait (i.e. postfire seedling) previously absent or weakly represented (Lamont *et al.*, 2013).

Charnov & Schaffer (1973) proposed a simple model to explain the evolutionary tradeoffs between annual and perennial life histories, more broadly characterized as semelparity versus iteroparity. Their model contends that life history evolution is driven by differences in adult versus offspring survivorship. The evolution of the perennial life history is expected when the average clutch size of an annual organism is increased by P : C individuals, where P and C are adult and juvenile survivorship, respectively. Thus, when adult survivorship through resprouting is high following a disturbance, and seedling survivorship is relatively low, resprouting should be of greater selective value than seedling recruitment. The seeding strategy is expected when resprouting success is low and seedling success is high (Fig. 1). The advantage of this model is that its multivariate nature allows factors affecting P and C to vary under different environmental conditions (e.g. Fig. 2), and thus under different temporal and spatial settings. This approach to understanding the relative roles of resprouting and seeding is more mechanistic, and more directly linked to key processes that enhance fitness, than previous models based on fire regime gradients.

In developing our model we concentrate on fire because it is a very widespread disturbance agent and has played a key role in plant evolution (Keeley & Rundel, 2005; Pausas & Keeley, 2009; Bond & Scott, 2010; Keeley *et al.*, 2011). Phylogenetic studies provide evidence of fire adaptation appearing at least since the early Paleocene (Crisp *et al.*, 2011; He *et al.*, 2012), and recent micro-evolutionary studies provide evidence of fire as an ongoing evolutionary pressure generating phenotypic divergence among populations (Gómez-González *et al.*, 2011; Pausas *et al.*, 2012; Hernández-Serrano *et al.*, 2013, 2014). All of these studies point to the increasing recognition of fire as an evolutionary force in plants (Pausas & Schwilk, 2012). In addition, fires currently occur in most regions world-wide and are susceptible to global change drivers (Pausas & Ribeiro, 2013; Pausas & Keeley, 2014); thus, fire regimes are changing world-wide. Understanding mechanisms of persistence under recurrent fires is of paramount importance for interpreting the past and predicting the future of our biota.

In the context of fire-prone ecosystems, postfire resprouting and postfire seeding are considered as two independent traits; that is, species living in those ecosystems may have one of the two traits, both, or none (Pausas *et al.*, 2004; see Box 1). This contrasts with other models in which resprouting and seeding are considered as alternative mechanisms of response to disturbance (Bellingham & Sparrow, 2000). One misconception arising from such approaches is to equate resprouting species with nonseeding species. All resprouters produce seeds; however, seeds of obligate

event per fire cycle and mostly with nonoverlapping generations. In fact, there is evidence of selection acting to favor early emergence in postfire seeding species (de Luis *et al.*, 2008). The two life histories have consequences in the cohort structure (obligate seeding species

Table 1 Main differences in life history processes among the three postfire strategies considered: species with one regeneration mechanism (resprouting or seeding) are called obligate postfire resprouters and obligate postfire seeders, respectively; species with both mechanisms are facultative postfire species (Fig. 3; Box 1)

	Obligate resprouters	Facultative seeders	Obligate seeders
Postfire resprouting	Yes	Yes	No
Postfire seeding	No	Yes	Yes
Life cycle	Polypyrlic (perennial)	Polypyrlic (perennial)	Monopyrlic (annual)
Reproduction cycles	Iteroparous ¹	Iteroparous ¹	Semelparous
Generations	Long and overlapping	Long and overlapping	Short and nonoverlapping
Recruitment	Between fires (fire-independent)	After fire (fire-dependent)	After fire (fire-dependent)
Synchrony of recruitment	Mainly asynchronous ²	Synchronous	Synchronous
Age structure of the population	Multiple cohort ²	Multiple cohort	Single cohort (even-aged)

We propose the terms 'polypyrlic' and 'monopyrlic' life cycles to be analogous to the perennial and annual life cycles but related to fire cycles instead of annual cycles (see main text and Box 1 for details).

¹Resprouters are commonly iteroparous but there are rare exceptions, such as *Agave* spp., that are semelparous (Schaffer & Schaffer, 1977). There are also some bamboo species that are clonal and semelparous but have a monopyrlic life cycle (Keeley & Bond, 1999).

²Obligate resprouters with strong mast-flowering may show synchronous recruitment and some even-aged cohorts, but their recruitment is still independent of fire.

resprouters are short-lived and establish only during interfire intervals, while facultative seeders regenerate by resprouting and reseed following fires (Box 1; Keeley, 1998). While in some ecosystems resprouting and postfire seeding may show a negative evolutionary correlation (i.e. the loss of one trait is evolutionarily linked to the acquisition of the other; Pausas & Verdú, 2005), it cannot be assumed as universal (Bond & Midgley, 2003; Pausas *et al.*, 2006). In fact the process of losing the resprouting capacity

is not physiologically (mechanistically) linked to the process of acquiring the capacity of postfire seeding, and thus these two processes require different explanations, even if the final outcome (extant functional strategies) implies a correlation. We suggest that changes in evolutionary pressures that modify adult (*P*) and juvenile (*C*) survival in postfire conditions determine the long-term success of each of the two regeneration mechanisms (Fig. 3). Specifically, we propose the following three hypotheses: (1) resprouting appeared early in plant evolution as a response to disturbance, and fire was an important driver in many lineages; (2) postfire seeding evolved under conditions where fires were

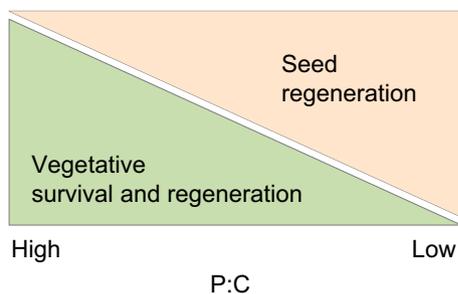


Fig. 1 The relative roles of vegetative and seed regeneration depend on the environmental pressures affecting the adult-to-offspring survival ratio (*P* : *C*).

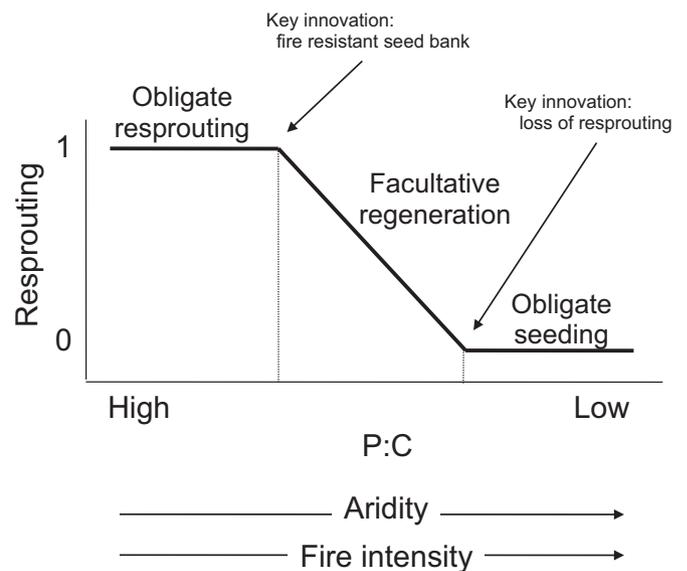


Fig. 3 Model of Fig. 1 applied to fire-prone ecosystems. Changes in the probability of resprouting along an adult-to-offspring survival ratio (*P* : *C*) gradient are not linear but show two turning points related to the acquisition of key innovations: the capacity to store a fire-resistant seed bank (postfire seeding), and the loss of resprouting capacity. Changes in the *P* : *C* ratio may be produced by different drivers (Fig. 2) and may have driven the rise of innovations during evolution.

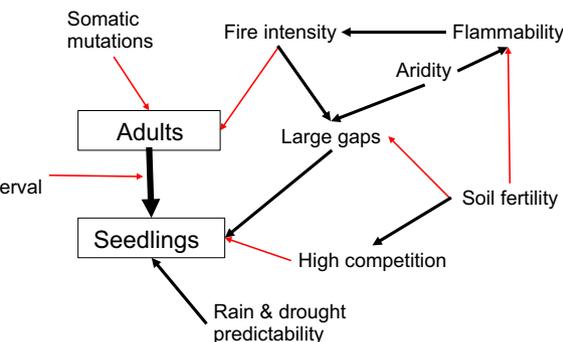


Fig. 2 Main factors affecting adult (*P*) and offspring seedling (*C*) survival, and thus the *P* : *C* ratio, in fire-prone ecosystems. Black arrows, positive effect; red arrows, negative effect.

predictable within the life span of the dominant plants and created conditions unfavorable for resprouting; and (3) the intensification of conditions favoring juvenile survival (C) and adult mortality (P) drove the loss of resprouting ability with the consequence of obligate-seeding species becoming entirely dependent on fire to complete their life cycle, with one generation per fire interval (monopyric life cycle).

Resprouting: an ancient and widespread trait

We define resprouting as the initiation of new shoots, usually from existing plant meristems, following fire or other disturbances that affect the whole plant. Basal resprouting is a nearly universal trait in perennial dicotyledonous plants (Wells, 1969), although in many broad-leaf trees it might be restricted to the sapling stage (Del Tredici, 2001; Shibata *et al.*, 2014). Our analysis of the 2741 species (in 951 genera and 139 families) living in Mediterranean ecosystems in different continents (data from Montenegro *et al.*, 2003; Pausas *et al.*, 2004, 2006; Pausas & Bradstock, 2007; Paula *et al.*, 2009) suggested that 57% have the ability to resprout, and that these resprouting species are distributed in 68% of the genera and in 90% of the families. The widespread taxonomical and phylogenetic distribution of resprouting ability and its presence in old lineages such as in Mesozoic gymnosperms (e.g. *Ephedra*, *Cycas*, *Wollemia nobilis*, *Ginkgo biloba* and *Sequoia sempervirens*; Pausas & Keeley, 2009), and in basal angiosperms (Feild *et al.*, 2004) suggest that it is an ancient trait in woody plants (Keeley *et al.*, 2012, chapter 9).

Plants resprout following many types of disturbance, and although of apparently adaptive value in fire-prone landscapes, resprouting is common in many vegetation types where fires are rare, such as rainforests or cold temperate ecosystems (Putz & Brokaw, 1989; Kauffman, 1991; Dietze & Clark, 2008; Shibata *et al.*, 2014). Given the long history of fire on Earth (Pausas & Keeley, 2009), it is plausible that in some lineages the evolution of resprouting may have been driven by fire, although this is not likely true for all lineages. Resprouting is not a simple trait and there are several mechanisms (Clarke *et al.*, 2013) that may be tied to different evolutionary frameworks. For instance, while epicormic resprouting can be clearly linked to fire (Crisp *et al.*, 2011), resprouting from roots, rhizomes, or even the root collar is not necessarily linked to fire (Lacey & Johnston, 1990). Resprouting from buds located in specialized and ontogenic structures such as lignotubers is also tied to fire-prone environments (Canadell & Zedler, 1995; Keeley *et al.*, 2012 chapter 3). In addition, some fire regimes may select against resprouting (see 'Loss of resprouting ability' below). Thus, resprouting is a complex functional trait that requires a careful analysis before its origin can be linked to a specific evolutionary pressure (Keeley *et al.*, 2011).

In moist and fertile environments, post-disturbance regeneration is very rapid and resprouting confers a competitive advantage by recapturing space previously occupied by the mature plant. This is because the surviving biomass (often belowground) enables the plant to quickly regenerate the aboveground biomass. In such conditions, seedlings fare poorly in competition with resprouts (Tyler & D'Antonio, 1995), and

thus the sexual reproduction of obligate resprouters is not usually linked to local disturbance, but rather these resprouters focus on strategies for finding appropriate microsites for recruitment (efficient dispersal methods, e.g. vertebrate dispersal). Consequently, such conditions do not select for delayed germination that restricts recruitment to postfire conditions (Keeley, 1998; Keeley *et al.*, 2012).

Given that resprouting is spatially, ecologically, and phylogenetically widespread, a reasonable hypothesis is that it is the ancestral state in most lineages, and that losing resprouting is a derived state (Wells, 1969; Bond & Midgley, 2003; Verdaguer & Ojeda, 2005). This does not discount the possibility of reversals in some lineages where occasionally resprouting species may originate from nonresprouting ancestors (Bond & Midgley, 2003; He *et al.*, 2011; Keeley *et al.*, 2012). In fire-prone ecosystems, a prerequisite for losing resprouting is to have an alternative postfire regeneration strategy, namely postfire seeding (Box 1). Thus, the appearance of postfire seeding should precede (or occur simultaneously with) the loss of resprouting capacity. Consequently, we hypothesize that facultative seeders were derived from resprouters and that nonresprouting obligate seeders were derived from facultative seeders. The loss and acquisition of these traits generate different trait combinations (regeneration strategies; Box 1; Fig. 3) that have very different population dynamics (Pausas, 1999) with strong ecological and evolutionary consequences (Table 1).

The acquisition of postfire seeding

Resprouting is optimal in fertile competitive environments where vigorous and aggressive resprouts rapidly occupy postfire gaps and therefore provide limited opportunities for seedling recruitment. By contrast, under stressful site conditions (such as low soil fertility and severe water deficits), fires may be more predictable, and the growth and the postfire resprouting rate are slower and the gaps with available resources are larger and longer lasting (Keeley & Zedler, 1978; Meentemeyer & Moody, 2002). Such conditions provide a substantial opportunity for recruitment leading to a decrease in P : C (Fig. 2) and selection for delaying reproduction to a single point in time (i.e. postfire) when more resources are available. By delaying germination to when conditions are optimal, plants also reduce fitness variance across fire cycles. Because fitness is a multiplicative process, it is very sensitive to occasional low values, and thus low fitness variance is selected for – despite the possible cost of increased seed mortality in the seed bank (Childs *et al.*, 2010). Factors that reduce survivorship of resprouting species would also create gaps; for example, fire regimes with recurrent but relatively long fire return intervals could contribute to selection for postfire gaps by accumulating fuels capable of killing resprouters by high-intensity fires (Moreno & Oechel, 1993; Lloret & López-Soria, 1993; Fig. 2). In addition, old-age mortality of resprouters during long interfire periods would also create gaps and opportunities for seeding regeneration of species with long seedbank longevity (Keeley & Zedler, 1978).

Aridity may reduce survivorship of both seedlings and adults; however, restricting reproduction to postfire conditions can enhance seedling survivorship because of the reduced competition

after fire. In addition, increased temperatures in postfire gaps may also accelerate seedling emergence under low water potential (Stevens *et al.*, 2014). There is evidence that the increasing aridification in marginal sites throughout the Tertiary could have driven the rise of the nonresprouting obligate seeding subgenus *Cerastes* in the genus *Ceanothus* (Ackerly *et al.*, 2006; Keeley *et al.*, 2012). The aridification process increased in extent during the Quaternary, and may have increased the diversification of seeder species (Verdú & Pausas, 2013). There is evidence from extant species that obligate seeders are more tolerant of water stress and fluctuations in water availability than resprouters, and exhibit physiological and anatomical traits that favor recruitment in open sites under stressful soil-drought conditions (Box 2). The expansion of drought-prone landscapes during the late Tertiary (Keeley *et al.*, 2012) resulted in an increase in fire size, which could also have favored delayed reproduction (dispersal in time) rather than dispersal in space.

Delaying reproduction until the advent of postfire conditions requires the accumulation of a seed bank during the fire-free interval, and a germination cue to recruit after fire when resources are most available. In different lineages this has been accomplished by various mechanisms such as serotiny and fire-dependent germination (Box 3). A corollary of having postfire seedling recruitment is the increase in population turnover, which increases the possibility of easily acquiring genetic novelties and thus better tracking changes in the environment from generation to generation (Wells, 1969; Raven, 1973; Schwilk & Kerr, 2002). Additionally, seed banks are also a source of genetic variation and novelty, as the aging of seeds and the reduction in viability implies an accumulation of mutations (Levin, 1990). Because resprouters regenerate in place, they are wedded to their environment, and acquiring postfire seeding makes the species better suited for a changing environment. All these factors may explain the high species richness in lineages that have acquired this trait (e.g. Fig. 4).

In addition to delayed reproduction enforced by deep seed dormancy and germination stimulated by heat or smoke, traits expected to be selected in postfire seeders include early emergence and rapid seedling growth. These traits allow seedlings to quickly capture resources and become more competitive, and this can be adaptive for seedlings under crowded postfire conditions where recruitment is concentrated in a single point in time (de Luis *et al.*, 2008). Concomitant with the evolution of postfire seeding, one might also expect selection for traits that enhance flammability. Flammability-enhancing traits could be favored in individuals if the elevated flammability resulted in increased mortality of neighbors, and thus opened space for recruitment opportunities for the flammable individual's offspring (Bond & Midgley, 1995). Modeling studies have suggested different genetic mechanisms for the evolution of flammability (Kerr *et al.*, 1999; Schwilk & Kerr, 2002), and there is some field evidence that recurrent fires increase plant flammability (Pausas & Moreira, 2012; Pausas *et al.*, 2012; Moreira *et al.*, 2014). The correlation between the postfire seeding strategy and flammability across different species provides further evidence for this hypothesis (Keeley & Zedler, 1998; Schwilk & Ackerly, 2001; Saura-Mas *et al.*, 2010; He *et al.*, 2012).

Box 2 Costs of resprouting

Resprouting carries a cost of storing resources to maintain a bud bank and support rapid post-disturbance regrowth. For instance, resprouting populations have higher levels of carbohydrates than nonresprouting species or populations, even at the seedling stage (Pate *et al.*, 1990; Verdaguer & Ojeda, 2002; Schwilk & Ackerly, 2005). This early allocation to buds might also imply reduced seedling growth. Repeated disturbance causes a reduction in root carbohydrates in resprouting species (Canadell *et al.*, 1991), and this reduction limits post-disturbance resprouting and increases mortality (Moreira *et al.*, 2012). To accumulate these resources, seedlings of resprouting plants allocate more resources to belowground storage while seedlings of nonresprouting plants allocate resources mainly to (aboveground) growth and reproduction (Bowen & Pate, 1993; Schutz *et al.*, 2009). The implications of this differential carbon allocation pattern are many, including a higher root : shoot ratio (Pate *et al.*, 1990; Verdaguer & Ojeda, 2002; Schwilk & Ackerly, 2005), lower seedling height growth (Pausas *et al.*, 2004), lower specific root length (Paula & Pausas, 2011), lower resistance to xylem cavitation (Jacobsen *et al.*, 2007; Pratt *et al.*, 2007, 2008; Vilagrosa *et al.*, 2013) and later maturity (Schwilk & Ackerly, 2005) in resprouting than in coexisting nonresprouting seeder species. In addition, seedlings of obligate seeders are often more efficient at conducting water (when it is available) than those of resprouters, and hence they are better adapted to take full advantage of periods with good water availability (Pratt *et al.*, 2010; Hernández *et al.*, 2011; Vilagrosa *et al.*, 2013). An additional implication is the potential differential response to anthropogenic changes in atmospheric CO₂ concentration (Bond & Midgley, 2012). At the landscape scale, the consequence of the differential carbon allocation pattern between postfire strategies is that resprouters tend to dominate in sites with more reliable water (pole-facing slopes, gullies, etc.) while seeders are able to occupy sites with stronger oscillations in water availability (Keeley, 1977; Pausas *et al.*, 1999; Clarke & Knox, 2002; Meentemeyer & Moody, 2002; Pausas & Bradstock, 2007; Coca & Pausas, 2012). Resource allocation models also predict higher storage capacity in high-productivity habitats (Iwasa & Kubo, 1997), and a lack of resprouting when disturbance is high in relation to productivity (Fig. Box 2, below). Despite the high carbon demand of resprouters, it is an omnipresent strategy in plants living in recurrently disturbed environments world-wide; and only in certain conditions in which safe sites are relatively frequent and large, and post-disturbance conditions reliable and predictable (Fig. 2), has the loss of this characteristic been successful (see main text).

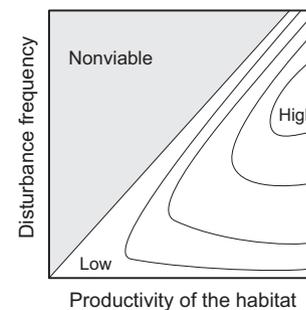


Fig. Box 2 Contours of the ratio of storage to growth allocation for an optimally growing resprouter in the space defined by habitat productivity and disturbance frequency (based on Iwasa & Kubo, 1997).

In our analogy of fire cycles being equivalent to annual cycles, and the loss of resprouting being analogous to a transition from the perennial to annual life history, we have considered the important

role of a decrease in P : C (Charnov & Schaffer, 1973). Semelparity is also favored by a low ratio of maturity age to time-between-reproductive-episodes (Young, 1981), which in our case is the ratio of maturity age to fire interval (Fig. 2; Box 3). Simulation models suggest that the most sensitive component in this process is seedling survival and the subsequent successful establishment as a mature reproductive adult (Bond & Van Wilgen, 1996; Ojeda *et al.*, 2005). Thus, factors contributing to seedling success, such as adaptation to arid sites enhanced by greater numbers of sexual generations, or allocation of resources to growth over storage, may be critical to the evolution of postfire seeding.

Loss of resprouting ability

Trait loss is a common evolutionary phenomenon when environmental change removes or weakens a source of selection that was previously important in maintaining a trait (Lathi *et al.*, 2009). In fire-prone shrublands, increased aridity during the evolutionary history provided the setting for the loss of resprouting in many lineages and for depending entirely on a single postfire pulse of seedlings, and thus becoming monopyric (obligate seeders; Box 1; Fig. 3). Despite resprouting being a good strategy for persisting, it carries costs, such as those related to storing carbohydrates, thickening the bark (in epicormic resprouting species), maintaining the bud bank for postfire regeneration and the associated reduction of both above- and belowground growth and delayed reproduction (Box 2). In addition, there are specific conditions linked to aridity in which adults are negatively affected (e.g. high-intensity fires and postfire droughts) and seedlings are promoted (e.g. increased availability of safe sites; Fig. 2), which facilitate the transition towards obligate seeding (lower P : C). Increasing aridity, however, makes postfire seedling recruitment more precarious and has selected for both physiological and anatomical traits that enhance drought tolerance (Box 2).

However, for seedling recruitment to become fully independent of resprouting, there must be a predictable fire regime (i.e. fire return intervals coupled with maturity age, plant longevity, and seed bank longevity; Box 3) and reliable postfire conditions (seasonal climates). While fire return intervals must fall within the life span of the plant (including the seedbank; Box 3), very short fire intervals preclude regeneration of obligate seeders (immaturity risk; Box 3), which explains why this transition has rarely occurred in savannas (fire intervals *c.* 1–5 yr). Examples of loss of resprouting are observed in the different Mediterranean-type ecosystems (where fire intervals vary from decades to centuries), such as the *Cerastes* subgenus of *Ceanothus* in North America, the genus *Cistus* in southern Europe, or the genus *Erica* in South Africa, among others (Fig. 4; Keeley *et al.*, 2012). In addition, there is evidence from extant species of such loss of resprouting resulting from the suppression of bud development responsible for lignotuber formation (Verdaguer & Ojeda, 2005). In fact, factors driving this transition could have been just an intensification of those responsible for the initial acquisition of postfire seeding in resprouters; so the loss of resprouting may have been occurring together with the optimization of the postfire seeding strategy, and factors working against resprouting may have favored postfire

Box 3 Postfire seeding: mechanisms of storing seeds and delaying germination

There are two modes of seed storage in postfire seeders: soil-stored seed banks and canopy-stored seed banks (also called serotiny). **Canopy-stored** seeds remain quiescent in closed woody structures (serotinous cones) for several to many years and seeds are released when the heat of the fire causes cones to open and disperse the seeds into the postfire soil bed (Lamont *et al.*, 1991). **Soil-stored** seeds remain dormant in the soil for decades to centuries and the heat of the fire or the chemicals from the combustion of organic matter break seed dormancy and stimulate germination after fire (i.e. heat-stimulated and smoke-stimulated germination; Keeley & Fotheringham, 2000). These different mechanisms of delaying reproduction correspond to different solutions for a similar 'problem', and may depend on phylogenetic constraints and the availability of pre-existing structures in the corresponding lineages on which natural selection could act. In addition, low-fertility soils may have selected against soil storage in favor of aerial storage because in such ecosystems high-nutrient seeds are subject to more intense predation when exposed on the soil surface (Keeley *et al.*, 2011).

Additionally, different fire regimes may play a critical role as soil-stored seeds can persist under long fire-free intervals whereas serotinous species cannot and thus such species require a *more predictable fire regime* (Fig. Box 3 below; Lamont *et al.*, 1991; Enright *et al.*, 1998; Lamont & Enright, 2000). Although serotinous cones may open with time or when the plant dies, the chance of recruiting without fire is low because of competition with existing vegetation, and when fire occurs the dead individual will not contribute to postfire seedling populations. By contrast, this is not necessarily true in soil seed bank species because seeds may remain in the soil for many years after the death of the parent plant waiting for the appropriate opportunity to germinate. Consequently, plants with soil-stored seed banks are more capable of dealing with higher variability in fire intervals than serotinous plants that require a more predictable fire regime.

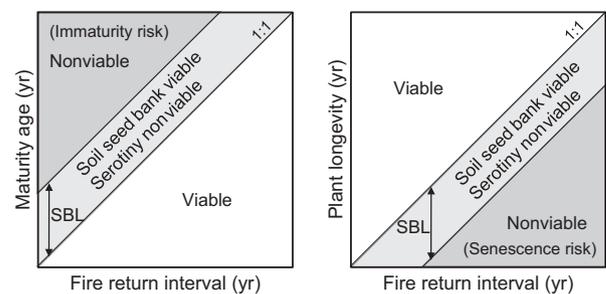


Fig. Box 3 The persistence of obligate postfire seeders is constrained by maturity age (left), plant longevity (right), and seed bank longevity (SBL), in relation to the fire-return interval, and these constrictions are stronger for serotinous seeders than for species with soil seed banks. Obligate postfire seeders are viable when the ratio of maturity age to fire interval is low and the ratio of longevity to fire interval is high. In soil seed bank seeders, these constraints are relaxed by the seed bank longevity. Nonviable conditions are also referred to as 'immaturity risk' (Zedler, 1995) where fire intervals are shorter than the maturity age plus the longevity of the soil seedbank, and 'senescence risk', where fire intervals are longer than plant longevity plus the longevity of the soil seedbank.

seeding. For instance, high-intensity fires negatively affect resprouting by killing vegetative buds (Lloret & López-Soria, 1993; Moreno & Oechel, 1993; Vesik *et al.*, 2004), and postfire droughts

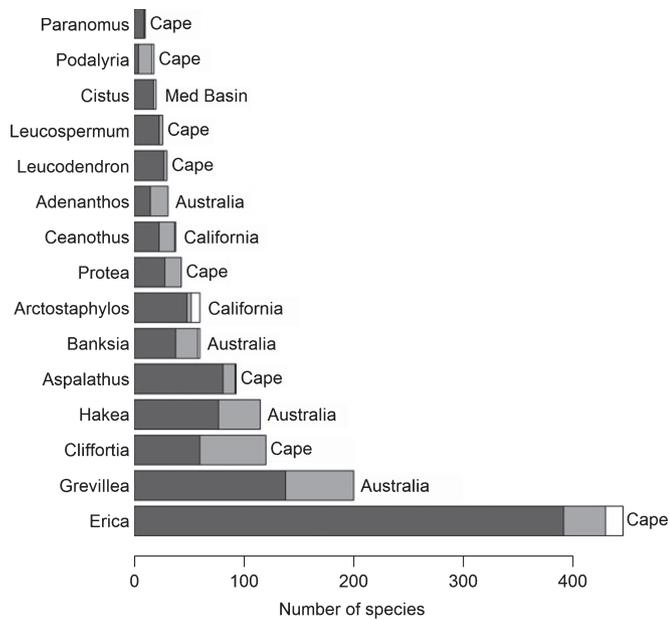


Fig. 4 Examples of spectacular radiations (high number of species per genus) for 15 genera (y-axis) living in fire-prone ecosystems of Australia, California, the Cape Region of South Africa, and the Mediterranean Basin. The number of postfire obligate (nonresprouting) seeders (dark gray), the number of postfire resprouters (light gray; mainly facultative species; see Box 1), and the number of species with variability in resprouting ability among populations (white) are shown. Data were compiled by Keeley *et al.* (2012).

reduce resprout success (Pratt *et al.*, 2014); but the increased intensity of fires might have also increased seed resistance to fire and the temperature thresholds for dormancy release (Moreira & Pausas, 2012). Thus, fire intensity favors postfire seeders by simultaneously increasing *C* and decreasing *P* (Figs 2, 3).

Another factor against resprouting is that genets may become very old and the potential exists for greater accumulation of deleterious alleles from somatic mutations (Wiens *et al.*, 1987). A high genetic load would potentially result in a reduced seed set (Lamont & Wiens, 2003). This hypothesis could explain the higher seed abortion rate in resprouting than in non-resprouting monocots in Western Australia (Meney *et al.*, 1997), and the anomalously nearly nonexistent seedling recruitment in some resprouters in Cape fynbos (*Retzia* spp.), *Banksia elegans* in Western Australian woodlands (Bond & Midgley, 2003), and *Adenostoma sparsifolium* in California (Wiens *et al.*, 2012). It could also explain the spatial genetic structure of some long-lived resprouters (Premoli & Steinke, 2008).

Selection of the obligate seeder life history would have been enhanced by their short generation time and rapid population turnover (under recurrent fires), which increases recombination-based mutations. In addition, the fact that seeders recruit from an aged seed bank also ensures the generation of genetic novelties (Levin, 1990). All of these processes increase the opportunity for natural selection to act, and thus speed up life history evolution (Wells, 1969; Schwilk & Kerr, 2002). The monopyric life cycle of obligate seeders precludes generational overlapping, which also contributes to increased genetic differentiation among populations and thus enhances evolutionary changes (Wade & McCauley, 1988; Nunney, 1993; Ellner & Hairston, 1994). These fast evolutionary changes should enable the species to better track

changes in the environment from generation to generation and adapt to micro-environmental conditions (Wells, 1969; Raven, 1973; Schwilk & Kerr, 2002), with the possible consequence of increasing diversification. In fact, the richness of many fire-prone ecosystems has been explained by the high diversification rates associated with the high population turnover in seeder species (Cowling & Pressey, 2001; Wisheu *et al.*, 2000; Barraclough, 2006; Fig. 4), as short generation times are related to high rates of molecular change (Smith & Donoghue, 2008). In this framework, there is some evidence of greater genetic diversity and differentiation in seeder than in resprouter populations of the same species (Segarra-Moragues & Ojeda, 2010), which can be considered an initial step to speciation (Coyne & Orr, 2004). However, diversification studies are still ambiguous in demonstrating a higher diversification of seeders, as other factors such as soil type and spatial heterogeneity may mask the pattern (Verdú *et al.*, 2007; Schnitzler *et al.*, 2011; Litsios *et al.*, 2014), and a broader analysis, including the three fire strategies (Table 1), remains to be carried out.

A successful coexistence

By losing the resprouting ability and acquiring the postfire seeding strategy, plants greatly increase their fitness in ecosystems with predictable fire recurrence, and thus access new ecological conditions causing rapid, and sometimes spectacular, adaptive radiations (Cowling & Pressey, 2001; Fig. 4). These novelties appear in different clades through convergent or parallel evolution (homoplastic novelties; Hunter, 1998) and have been highly successful in ecological and evolutionary terms. Thus, they could be considered as key evolutionary innovations (Hunter, 1998) in crown-fire ecosystems. In fact, most fire-prone ecosystems are included in the list of global biodiversity hotspots (Myers *et al.*, 2000) and the evolutionary pathways shaping fire-related traits may contribute to their diversity. A consequence of the proposed model is that, in lineages that have acquired postfire seeding, this trait is more phylogenetically conserved than resprouting, as has been observed in different Proteaceae lineages (Bond & Midgley, 2003; He *et al.*, 2011) and in *Ceanothus* (Ackerly, 2003).

We have described the main transitions in fire traits driven by the historical increase in aridity and the concomitant increase in the predictable role of fire in some landscapes (Figs 2, 3). Although the primary factors driving these transitions are likely to be similar in fire-prone environments, the timing of the evolution of each of the novelties has varied across different regions and lineages. For example, the widespread and very old infertile soils in southwestern Australia would have assembled flammable vegetation early in the Tertiary (He *et al.*, 2011, 2012; Crisp *et al.*, 2011; Keeley *et al.*, 2012; chapter 10), whereas in other regions, aridity may have played a larger role in diversification at a later point (Verdú & Pausas, 2013). In addition, not all plant lineages were subject to the same transitions because the ecological and evolutionary forces that shaped plants changed in time and space (see box 1 from Keeley *et al.*, 2011), and plants in different regions had different phylogenetic (historic) constraints. Consequently, alternative pathways also exist (Box 4) and thus differing strategies may coexist.

Box 4 Evolutionary transitions

If we note the two traits (resprouting and postfire seeding) as R and P, and the two states of each trait as + and –, the four possible combinations define the four general postfire strategies (Box 1 and Pausas *et al.*, 2004): R+P– (obligate resprouters); R+P+ (facultative species); R–P+ (obligate seeders); and R–P– (without endogenous regeneration, postfire colonizers). By changing the state of one of the traits, species may evolve from one strategy to another. Fig. Box 4 (below) shows some possible evolutionary transitions. For instance (Fig. Box 4 below, clockwise), the acquisition of postfire seeding by obligate resprouters (R+P–, ancestral strategy) might have occurred in *Arctostaphylos* and *Ceanothus* species (Keeley *et al.*, 2012), leading to some facultative species (R+P+) in these genera (Fig. 4). The reversal may have occurred in the Chilean matorral where fire was probably frequent before the rise of the Andes that currently limit the summer storms (Keeley *et al.*, 2012). Loss and acquisition of resprouting in seeders are common in Proteaceae (Bond & Midgley, 2003; Lamont *et al.*, 2013) and Fabaceae (Boatwright *et al.*, 2008); and aberrant resprouting individuals of obligate seeders have been observed in Lamiaceae in the Mediterranean basin (J. G. Pausas, personal observations), in Ericaceae and Fabaceae in the South African Cape region (Schutte *et al.*, 1995; Ojeda, 1998) and in the Australian *Banksia* and *Hakea* (R. Bradstock, pers. comm.). The acquisition of postfire seeding by nonresprouters can be exemplified by the acquisition of serotiny in *Pinus* during the Cretaceous (He *et al.*, 2012). Loss and acquisition of resprouting from the ancestral obligate resprouter seem to have occurred in conifers (*Juniperus* and *Pinus*; He *et al.*, 2012).

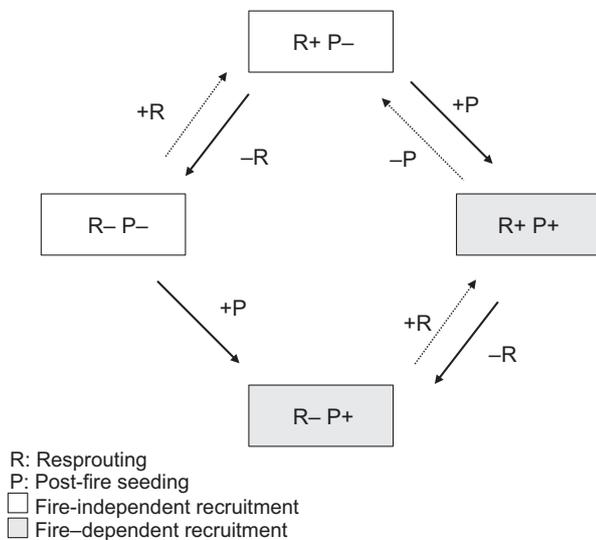


Fig. Box 4 Possible evolutionary transitions between postfire regeneration strategies. Continuous lines indicate more frequent transitions than dotted lines.

Considering that landscape heterogeneity leads to heterogeneous fire regimes, it is to be expected that evolution would select for diverse life history solutions to fire, including both polypyrnic and monopyrnic life cycles (i.e. resprouters and obligate seeders, respectively), and that communities would assemble that include this variety of strategies. In fact, having different regeneration

niches is one of the mechanisms of coexistence in disturbance-prone ecosystems (Grubb, 1977; Lavorel & Chesson, 1995; Miller & Chesson, 2009), and Mediterranean ecosystems provide examples of such coexistence (Pausas *et al.*, 2004; Keeley *et al.*, 2012; Marais *et al.*, 2014). Spatial models of iteroparity and semelparity dynamics also suggest long-term coexistence when life-history traits (number of offspring and juvenile and adult survival) are variable in time (Ranta *et al.*, 2002). In accordance with the evolutionary tendency of resprouting being lost with aridity and seeders being more capable to recruit in drier conditions (Box 2), there is evidence from many regions world-wide of some spatial landscape segregation between postfire seeders and obligate resprouters, in such a way that the former occupy the parts of the landscape where water is more susceptible to strong oscillations, and the latter occupy parts where water is more reliable (Box 2). The coexistence of species with different regeneration strategies reflects that different combinations of traits can be adaptive in fire-prone ecosystems (Box 4). This is an alternative view to the one considering that strategies with ancestral trait states (such as obligate resprouters) are relict and can only persist thanks to the facilitative effect of species with derived trait states (such as obligate seeders) (sensu Valiente-Banuet *et al.*, 2006). However, simulation studies of population dynamics in variable fire regimes find coexistence possible between any two of the three strategies, but not for all three (Cowan, 2010).

In conclusion, our models based on the relative survivorships of adults and juveniles coupled with the restrictions imposed by fire return intervals provide a useful framework for understanding temporal and spatial variations in resprouting and seeding in ecosystems subjected to crown-fire regimes. This model accounts for patterns of coexistence and environmental changes contributing to the evolution of seeding from resprouting ancestors.

Acknowledgements

This work has been performed under the framework of the TREVOL projects (CGL2012-39938-C02-01) from the Spanish government. Centro de Investigaciones sobre Desertificación (CIDE; Desertification Research Center) is a joint institute of the Spanish National Research Council (CSIC), the University of Valencia and Generalitat Valenciana. We thank the anonymous referees and D. Ackerly for helpful comments. The authors declare no conflict of interest. Any use of trade names is for descriptive purposes only and does not imply endorsement by the US Government.

References

- Ackerly D. 2003. Community assembly, niche conservatism, and adaptive evolution in changing environments. *International Journal of Plant Sciences* 164: 165–184.
- Ackerly DD, Schwillk DW, Webb CO. 2006. Niche evolution and adaptive radiation: testing the order of trait divergence. *Ecology* 87: s50–s61.
- Barraclough TG. 2006. What can phylogenetics tell us about speciation in the Cape flora? *Diversity and Distributions* 12: 21–26.
- Bellingham PJ, Sparrow AD. 2000. Resprouting as a life history strategy in woody plant communities. *Oikos* 89: 409–416.
- Boatwright JS, Savolainen V, Van Wyk BE, Schutte-Vlok L, Forest F, Van Der Bank M. 2008. Systematic position of the anomalous genus *Cadia* and the phylogeny of the tribe Podalyrieae (Fabaceae). *Systematic Botany* 33: 133–147.

- Bond WJ, Midgley JJ. 1995. Kill thy neighbour: an individualistic argument for the evolution of flammability. *Oikos* 73: 79–85.
- Bond WJ, Midgley JJ. 2001. Ecology of sprouting in woody plants: the persistence niche. *Trends in Ecology and Evolution* 16: 45–51.
- Bond WJ, Midgley JJ. 2003. The evolutionary ecology of sprouting in woody plants. *International Journal of Plant Sciences* 164: 103–114.
- Bond WJ, Midgley GF. 2012. Carbon dioxide and the uneasy interactions of trees and savannah grasses. *Philosophical Transactions of the Royal Society of London, Series B: Biological Sciences* 367: 601–612.
- Bond WJ, Scott AC. 2010. Fire and the spread of flowering plants in the Cretaceous. *New Phytologist* 188: 1137–1150.
- Bond WJ, Van Wilgen BW. 1996. *Fire and plants*. London, UK: Chapman & Hall.
- Bowen BJ, Pate JS. 1993. The significance of root starch in post-fire shoot recovery of the resprouter *Stirlingia latifolia* R. Br. (Proteaceae). *Annals of Botany* 72: 7–16.
- Canadell J, Lloret F, López-Soria L. 1991. Resprouting vigour of two Mediterranean shrub species after experimental fire treatments. *Vegetatio* 95: 119–126.
- Canadell J, Zedler PH. 1995. Underground structures of woody plants in Mediterranean ecosystems of Australia, California, and Chile. In: Arroyo MTK, Zedler PH, Fox MD, eds. *Ecology and biogeography of Mediterranean ecosystems in Chile, California and Australia*. New York, NY, USA: Springer-Verlag, 177–210.
- Charnov EL, Schaffer WM. 1973. Life-history consequences of natural selection: Cole's result revisited. *American Naturalist* 107: 791.
- Childs DZ, Metcalf CJE, Rees M. 2010. Evolutionary bet-hedging in the real world: empirical evidence and challenges revealed by plants. *Proceedings of the Royal Society B: Biological Sciences* 277: 3055–3064.
- Clarke PJ, Knox JE. 2002. Post-fire response of shrubs in the tablelands of eastern Australia: do existing models explain habitat differences? *Australian Journal of Botany* 50: 53–62.
- Clarke PJ, Lawes MJ, Midgley JJ, Lamont BB, Ojeda F, Burrows GE, Enright NJ, Knox KJE. 2013. Resprouting as a key functional trait: how buds, protection and resources drive persistence after fire. *New Phytologist* 197: 19–35.
- Coca M, Pausas JG. 2012. Scale-dependent segregation of seeders and resprouters in cork oak (*Quercus suber*) forests. *Oecologia* 168: 503–510.
- Cowan PD. 2010. *Flammability, physiology and coexistence in fire prone environments*. Ph.D. thesis, University of California, Berkeley, CA, USA.
- Cowling RM, Pressey RM. 2001. Rapid plant diversification: planning for an evolutionary future. *Proceedings of the National Academy of Sciences, USA* 98: 5452–5457.
- Coyne JA, Orr HA. 2004. *Speciation*. Sunderland, MA, USA: Sinauer Associates.
- Crisp MD, Burrows GE, Cook LG, Thornhill AH, Bowman DMJS. 2011. Flammable biomes dominated by eucalypts originated at the Cretaceous-Palaeogene boundary. *Nature Communications* 2: 193.
- Del Tredici P. 2001. Sprouting in temperate trees: a morphological and ecological review. *The Botanical Review* 67: 121–140.
- Dietze MC, Clark JS. 2008. Changing the gap dynamics paradigm: vegetative regeneration control on forest response to disturbance. *Ecological Monographs* 78: 331–347.
- Ellner S, Hairston NG. 1994. Role of overlapping generations in maintaining genetic-variation in a fluctuating environment. *American Naturalist* 143: 403–417.
- Enright NJ, Marsula R, Lamont BB, Wissel C. 1998. The ecological significance of canopy seed storage in fire-prone environments: a model for non-resprouting shrubs. *Journal of Ecology* 86: 946–959.
- Feild TS, Arens NC, Doyle JA, Dawson TE, Donoghue MJ. 2004. Dark and disturbed: a new image of early angiosperm ecology. *Paleobiology* 30: 82.
- Gómez-González S, Torres-Díaz C, Bustos-Schindler C, Gianoli E. 2011. Anthropogenic fire drives the evolution of seed traits. *Proceedings of the National Academy of Sciences, USA* 108: 18743–18747.
- Grubb PJ. 1977. The maintenance of species-richness in plant communities: the importance of the regeneration niche. *Biological Reviews* 52: 107–145.
- He T, Lamont BB, Downes KS. 2011. *Banksia* born to burn. *New Phytologist* 191: 184–196.
- He T, Pausas JG, Belcher CM, Schwillk DW, Lamont BB. 2012. Fire-adapted traits of *Pinus* arose in the fiery Cretaceous. *New Phytologist* 194: 751–759.
- Hernández EI, Pausas JG, Vilagrosa A. 2011. Leaf physiological traits in relation to resprouter ability in the Mediterranean Basin. *Plant Ecology* 212: 1959–1966.
- Hernández-Serrano A, Verdú M, González-Martínez SC, Pausas JG. 2013. Fire structures pine serotiny at different scales. *American Journal of Botany* 100: 2349–2356.
- Hernández-Serrano A, Verdú M, Santos-del-Blanco L, Climent J, González-Martínez SC, Pausas JG. 2014. Heritability and quantitative genetic divergence of serotiny, a fire persistence plant trait. *Annals of Botany*. doi: 10.1093/aob/mcu142.
- Hunter JP. 1998. Key innovations and the ecology of macroevolution. *Trends in Ecology & Evolution* 13: 31–36.
- Iwasa Y, Kubo T. 1997. Optimal size of storage for recovery after unpredictable disturbance. *Evolutionary Ecology* 11: 41–65.
- Jacobsen AL, Agenbag L, Esler KJ, Pratt RB, Ewers FW, Davis SD. 2007. Xylem density, biomechanics and anatomical traits correlate with water stress in 17 evergreen shrub species of the Mediterranean-type climate region of South Africa. *Ecology* 95: 171–183.
- Kauffman JB. 1991. Survival by sprouting following fire in tropical forests of the eastern Amazon. *Biotropica* 23: 219–224.
- Keeley JE. 1977. Seed production, seed populations in soil, and seedling production after fire for two congeneric pairs of sprouting and nonsprouting Chaparral shrubs. *Ecology* 58: 820–829.
- Keeley JE. 1986. Resilience of mediterranean shrub communities to fire. In: Dell B, Hopkins AJM, Lamont BB, eds. *Resilience in mediterranean-type ecosystems*. Dordrecht, the Netherlands: Dr W. Junk Publishers, 95–112.
- Keeley JE. 1998. Coupling demography, physiology and evolution in chaparral shrubs. In: Rundel PW, Montenegro G, Jaksic FM, eds. *Landscape disturbance and biodiversity in Mediterranean-type ecosystems*. Berlin, Germany: Springer, 257–264.
- Keeley JE, Bond WJ. 1999. Mast flowering and semelparity in bamboos: The bamboo fire cycle hypothesis. *American Naturalist* 154: 383–391.
- Keeley JE, Bond WJ, Bradstock RA, Pausas JG, Rundel PW. 2012. *Fire in Mediterranean ecosystems: ecology, evolution and management*. Cambridge, UK: Cambridge University Press.
- Keeley JE, Fotheringham CJ. 2000. Role of fire in regeneration from seeds. In: Fenner M, ed. *Seeds: the ecology of regeneration in plant communities*. Wallingford, UK: CAB International, 311–330.
- Keeley JE, Pausas JG, Rundel PW, Bond WJ, Bradstock RA. 2011. Fire as an evolutionary pressure shaping plant traits. *Trends in Plant Science* 16: 406–411.
- Keeley JE, Rundel PW. 2005. Fire and the Miocene expansion of C₄ grasslands. *Ecology Letters* 8: 683–690.
- Keeley JE, Zedler PH. 1978. Reproduction of chaparral shrubs after fire: a comparison of sprouting and seedling strategies. *American Midland Naturalist* 99: 142–161.
- Keeley JE, Zedler PH. 1998. Evolution of life histories in *Pinus*. In: Richardson DM, ed. *Ecology and biogeography of Pinus*. Cambridge, UK: Cambridge University Press, 219–250.
- Kerr B, Schwillk DW, Bergman A, Feldman MW. 1999. Rekindling an old flame: a haploid model for the evolution and impact of flammability in resprouting plants. *Evolutionary Ecology Research* 1: 807–833.
- Klimesová J, Klimes L. 2003. Resprouting of herbs in disturbed habitats: is it adequately described by Bellingham-Sparrow's model? *Oikos* 103: 225–229.
- Klimesová J, Klimes L. 2007. Bud banks and their role in vegetative regeneration—a literature review and proposal for simple classification and assessment. *Perspectives in Plant Ecology, Evolution and Systematics* 8: 115–129.
- Knox KJE, Clarke PJ. 2005. Nutrient availability induces contrasting allocation and starch formation in resprouting and obligate seeding shrubs. *Functional Ecology* 19: 690–698.
- Lacey C, Johnston R. 1990. Woody clumps and clumpwoods. *Australian Journal of Botany* 38: 299–334.
- Lahti DC, Johnson NA, Ajie BC, Otto SP, Hendry AP, Blumstein DT, Coss RG, Donohue K, Foster SA. 2009. Relaxed selection in the wild. *Trends in Ecology & Evolution* 24: 487–496.
- Lamont BB, Enright NJ. 2000. Adaptive advantages of aerial seed banks. *Plant Species Biology* 15: 157–166.
- Lamont BB, Enright NJ, He T. 2011. Fitness and evolution of resprouters in relation to fire. *Plant Ecology* 2012: 1945–1957.

- Lamont BB, He T, Downes K. 2013. Adaptive responses to directional trait selection in the Miocene enabled Cape proteas to colonize the savanna grasslands. *Evolutionary Ecology* 27: 1099–1115.
- Lamont BB, Le Maitre DC, Cowling RM, Enright NJ. 1991. Canopy seed storage in woody plants. *The Botanical Review* 57: 277–317.
- Lamont BB, Wiens D. 2003. Are seed set and speciation rates always low among species that resprout after fire, and why? *Evolutionary Ecology* 17: 277–292.
- Lavorel S, Chesson P. 1995. How species with different regeneration niches coexist in patchy habitats with local disturbances. *Oikos* 74: 103–114.
- Levin DA. 1990. The seed bank as a source of genetic novelty in plants. *American Naturalist* 135: 563–572.
- Litsios G, Wüest RO, Kostikova A, Forest F, Lexer C, Linder HP, Pearman PB, Zimmermann NE, Salamin N. 2014. Effects of a fire response trait on diversification in replicated radiations. *Evolution* 68: 453–465.
- Lloret F, López-Soria L. 1993. Resprouting of *Erica multiflora* after experimental fire treatments. *Journal of Vegetation Science* 4: 367–374.
- de Luis M, Verdú M, Raventós J. 2008. Early to rise makes a plant healthy, wealthy, and wise. *Ecology* 89: 3061–3071.
- Marais KE, Pratt RB, Jacobs SM, Jacobsen AL, Esler KJ. 2014. Postfire regeneration of resprouting mountain fynbos shrubs: differentiating obligate resprouters and facultative seeders. *Plant Ecology* 215: 195–208.
- Meentemeyer RK, Moody A. 2002. Distribution of plant life history types in California chaparral: the role of topographically-determined drought severity. *Journal of Vegetation Science* 13: 67–78.
- Meney KA, Dixon KW, Pate JS. 1997. Reproductive potential of obligate seeder and resprouter herbaceous perennial monocots (Restionaceae, Anarthriaceae, Ecdiaceae) from South-western Western Australia. *Australian Journal of Botany* 45: 771–782.
- Miller AD, Chesson P. 2009. Coexistence in disturbance-prone communities: how a resistance-resilience trade-off generates coexistence via the storage effect. *American Naturalist* 173: E30–E43.
- Montenegro G, Gómez M, Díaz F, Ginocchio R. 2003. Regeneration potential of Chilean matorral after fire: an updated view. In: Veblen TT, Baker WL, Montenegro G, Swetnam TW, eds. *Fire and climatic change in temperate ecosystems of the Western Americas*. New York, NY, USA: Springer-Verlag, 381–409.
- Moreira B, Castellanos MC, Pausas JG. 2014. Genetic component of flammability variation in a Mediterranean shrub. *Molecular Ecology* 23: 1213–1223.
- Moreira B, Pausas JG. 2012. Tanned or burned: the role of fire in shaping physical seed dormancy. *PLoS ONE* 7: e39810.
- Moreira B, Tormo J, Pausas JG. 2012. To resprout or not to resprout: factors driving intraspecific variability in resprouting. *Oikos* 121: 1577–1584.
- Moreno JM, Oechel WC. 1993. Demography of *Adenostoma fasciculatum* after fires of different intensities in Southern California Chaparral. *Oecologia* 96: 95–101.
- Myers N, Mittermeier RA, Mittermeier CG, da Fonseca GAB, Kent J. 2000. Biodiversity hotspots for conservation priorities. *Nature* 403: 853–858.
- Nathan R, Safriel UN, Noy-Meir I, Schiller G. 1999. Seed release without fire in *Pinus halepensis*, a Mediterranean serotinous wind-dispersed tree. *Journal of Ecology* 87: 659–669.
- Nunney L. 1993. The influence of mating system and overlapping generations on effective population size. *Evolution* 47: 1329–1341.
- Ojeda F. 1998. Biogeography of seeder and resprouter *Erica* species in the Cape Floristic Region – where are the resprouters? *Biological Journal of the Linnean Society* 63: 331–347.
- Ojeda F, Brun FG, Vergara JJ. 2005. Fire, rain and the selection of seeder and resprouter life-histories in fire-recruiting, woody plants. *New Phytologist* 168: 155–165.
- Pate JS, Froend RH, Bowen BJ, Hansen A, Kuo J. 1990. Seedling growth and storage characteristics of seeder and resprouter species of Mediterranean-type ecosystems of SW Australia. *Annals of Botany* 65: 585–601.
- Paula S, Arianoutsou M, Kazanis D, Tavsanoglu C, Lloret F, Buhk C, Ojeda F, Luna B, Moreno JM, Rodrigo A et al. 2009. Fire-related traits for plant species of the Mediterranean Basin. *Ecology* 90: 1420.
- Paula S, Pausas JG. 2011. Root traits explain different foraging strategies between resprouting abilities. *Oecologia* 165: 321–331.
- Pausas JG. 1999. Mediterranean vegetation dynamics: modelling problems and functional types. *Plant Ecology* 140: 27–39.
- Pausas JG. 2001. Resprouting vs seeding – a Mediterranean perspective. *Oikos* 94: 193–194.
- Pausas JG, Alessio GA, Moreira B, Corcobado G. 2012. Fires enhance flammability in *Ulex parviflorus*. *New Phytologist* 193: 18–23.
- Pausas JG, Bradstock RA. 2007. Fire persistence traits of plants along a productivity and disturbance gradient in Mediterranean shrublands of SE Australia. *Global Ecology and Biogeography* 16: 330–340.
- Pausas JG, Bradstock RA, Keith DA, Keeley JE, GCTE_Fire_Network. . 2004. Plant functional traits in relation to fire in crown-fire ecosystems. *Ecology* 85: 1085–1100.
- Pausas JG, Carbó E, Caturra RN, Gil JM, Vallejo R. 1999. Post-fire regeneration patterns in the eastern Iberian Peninsula. *Acta Oecologica* 20: 499–508.
- Pausas JG, Keeley JE. 2009. A burning story: the role of fire in the history of life. *BioScience* 59: 593–601.
- Pausas JG, Keeley JE. 2014. Abrupt climate-independent fire regime changes. *Ecosystems*. doi:10.1007/s10021-014-9773-5.
- Pausas JG, Keeley JE, Verdú M. 2006. Inferring differential evolutionary processes of plant persistence traits in Northern Hemisphere Mediterranean fire-prone ecosystems. *Journal of Ecology* 94: 31–39.
- Pausas JG, Moreira B. 2012. Flammability as a biological concept. *New Phytologist* 194: 610–613.
- Pausas JG, Ribeiro E. 2013. The global fire-productivity relationship. *Global Ecology and Biogeography* 22: 728–736.
- Pausas JG, Schilck DW. 2012. Fire and plant evolution. *New Phytologist* 193: 301–303.
- Pausas JG, Verdú M. 2005. Plant persistence traits in fire-prone ecosystems of the Mediterranean basin: a phylogenetic approach. *Oikos* 109: 196–202.
- Pratt R, Jacobsen A, Mohla R, Ewers F, Davis S. 2008. Linkage between water stress tolerance and life history type in seedlings of nine chaparral species (Rhamnaceae). *Journal of Ecology* 96: 1252–1265.
- Pratt RB, Jacobsen AL, Golgotiu KA, Sperry JS, Ewers FW, Davis SD. 2007. Life history type and water stress tolerance in nine California chaparral species (Rhamnaceae). *Ecological Monographs* 77: 239–253.
- Pratt RB, Jacobsen AL, Ramirez AR, Helms AM, Traugh CA, Tobin MF, Heffner MS, Davis SD. 2014. Mortality of resprouting chaparral shrubs after a fire and during a record drought: physiological mechanisms and demographic consequences. *Global Change Biology* 20: 893–907.
- Pratt RB, North GB, Jacobsen AL, Ewers FW, Davis SD. 2010. Xylem root and shoot hydraulics is linked to life history type in chaparral seedlings. *Functional Ecology* 24: 70–81.
- Premoli AC, Steinke L. 2008. Genetics of sprouting: effects of long-term persistence in fire-prone ecosystems. *Molecular Ecology* 17: 3827–3835.
- Putz FE, Brokaw NVL. 1989. Sprouting of broken trees on Barro Colorado Island, Panama. *Ecology* 70: 508–512.
- Ranta ESA, Tesar D, Kaitala V. 2002. Environmental variability and semelparity vs. iteroparity as life histories. *Journal of Theoretical Biology* 217: 391–396.
- Raven PH. 1973. The evolution of Mediterranean floras. In: di Castri F, Mooney HA, eds. *Mediterranean ecosystems: origin and structure*. New York, NY, USA: Springer-Verlag, 213–223.
- Saura-Mas S, Paula S, Pausas JG, Lloret F. 2010. Fuel loading and flammability in the Mediterranean Basin woody species with different post-fire regenerative strategies. *International Journal of Wildland Fire* 19: 783–794.
- Schaffer W, Schaffer M. 1977. The adaptive significance of variations in reproductive habit in the Agavaceae. *Evolutionary Ecology* 261: 1051–1069.
- Schnitzler J, Barraclough TG, Boatwright JS, Goldblatt P, Manning JC, Powell MP, Rebelo T, Savolainen V. 2011. Causes of plant diversification in the Cape biodiversity hotspot of South Africa. *Systematic Biology* 60: 343–357.
- Schutte AL, Vlok JHJ, Vanwyk BE. 1995. Fire-survival strategy – a character of taxonomic, ecological and evolutionary importance in fynbos legumes. *Plant Systematics & Evolution* 195: 243–259.
- Schutz AEN, Bond WJ, Cramer MD. 2009. Juggling carbon: allocation patterns of a dominant tree in a fire-prone savanna. *Oecologia* 160: 235–246.
- Schilck DW, Ackerly DD. 2001. Flammability and serotiny as strategies: correlated evolution in pines. *Oikos* 94: 326–336.
- Schilck DW, Ackerly DD. 2005. Is there a cost to resprouting? Seedling growth rate and drought tolerance in sprouting and nonsprouting *Ceanothus* (Rhamnaceae). *American Journal of Botany* 92: 404–410.
- Schilck DW, Kerr B. 2002. Genetic niche-hiking: an alternative explanation for the evolution of flammability. *Oikos* 99: 431–442.

- Segarra-Moragues JG, Ojeda F. 2010. Post-fire response and genetic diversity in *Erica coccinea*: connecting population dynamics and diversification in a biodiversity hotspot. *Evolution* 64: 3511–3524.
- Shibata R, Shibata M, Tanaka H, Iida S, Masaki T, Hatta F, Kurokawa H, Nakashizuka T. 2014. Interspecific variation in the size-dependent resprouting ability of temperate woody species and its adaptive significance. *Journal of Ecology* 102: 209–220.
- Smith SA, Donoghue MJ. 2008. Rates of molecular evolution are linked to life history in flowering plants. *Science* 322: 86–89.
- Stevens N, Seal C, Archibald S, Bond W. 2014. Increasing temperatures can improve seedling establishment in arid-adapted savanna trees. *Oecologia* 175: 1029–1040.
- Tyler CM, D'Antonio CM. 1995. The effects of neighbors on the growth and survival of shrubs seedling following fire. *Oecologia* 102: 255–264.
- Valiente-Banuet A, Rumebe AV, Verdú M, Callaway RM. 2006. Modern quaternary plant lineages promote diversity through facilitation of ancient tertiary lineages. *Proceedings of the National Academy of Sciences, USA* 103: 16812–16817.
- Verdguer D, Ojeda F. 2002. Root starch storage and allocation patterns in seeder and resprouter seedlings of two Cape *Erica* (Ericaceae) species. *American Journal of Botany* 89: 1189–1196.
- Verdguer D, Ojeda F. 2005. Evolutionary transition from resprouter to seeder life history in two *Erica* (Ericaceae) species: insights from seedling axillary buds. *Annals of Botany* 95: 593–599.
- Verdú M, Pausas JG. 2013. Syndrome-driven diversification in a Mediterranean ecosystem. *Evolution* 67: 1756–1766.
- Verdú M, Pausas JG, Segarra-Moragues JG, Ojeda F. 2007. Burning phylogenies: fire, molecular evolutionary rates, and diversification. *Evolution* 61: 2195–2204.
- Vesk P, Westoby M. 2004. Sprouting ability across diverse disturbances and vegetation types worldwide. *Journal of Ecology* 92: 310–320.
- Vesk PA, Warton DI, Westoby M. 2004. Sprouting by semi-arid plants: testing a dichotomy and predictive traits. *Oikos* 107: 72–89.
- Vilagrosa A, Hernández E, Luis VC, Cochar H, Pausas JG. 2013. Physiological differences explain the co-existence of different regeneration strategies in Mediterranean ecosystems. *New Phytologist* 201: 1277–1288.
- Wade MJ, McCauley DE. 1988. Extinction and recolonization: their effects on the genetic differentiation of local populations. *Evolution* 42: 995–1005.
- Wells PV. 1969. The relation between mode of reproduction and extent of specialization in woody genera of the California chaparral. *Evolution* 23: 264–267.
- Wiens D, Allphin L, Wall M, Slaton MR, Davis SD. 2012. Population decline in *Adenostoma sparsifolium* (Rosaceae): an ecogenetic hypothesis for background extinction. *Biological Journal of the Linnean Society* 105: 269–292.
- Wiens D, Calvin CL, Wilson CA, Davern CI, Frank D, Seavey SR. 1987. Reproductive success, spontaneous embryo abortion, and genetic load in flowering plants. *Oecologia* 71: 501–509.
- Wisheu IC, Rosenzweig ML, Olsvig-Whittaker L, Shmida A. 2000. What makes nutrient-poor mediterranean heathlands so rich in plant diversity? *Evolutionary Ecology Research* 2: 935–955.
- Young TP. 1981. A general model of comparative fecundity for semelparous and iteroparous life histories. *American Naturalist* 118: 27–36.
- Zedler PH. 1995. Fire frequency in southern California shrublands: biological effects and management options. In: Keeley JE, Scott T, eds. *Brushfires in California: ecology and management*. Fairfield, WA, USA: International Association of Wildland Fire, 101–112.



About New Phytologist

- *New Phytologist* is an electronic (online-only) journal owned by the New Phytologist Trust, a **not-for-profit organization** dedicated to the promotion of plant science, facilitating projects from symposia to free access for our Tansley reviews.
- Regular papers, Letters, Research reviews, Rapid reports and both Modelling/Theory and Methods papers are encouraged. We are committed to rapid processing, from online submission through to publication 'as ready' via *Early View* – our average time to decision is <25 days. There are **no page or colour charges** and a PDF version will be provided for each article.
- The journal is available online at Wiley Online Library. Visit **www.newphytologist.com** to search the articles and register for table of contents email alerts.
- If you have any questions, do get in touch with Central Office (np-centraloffice@lancaster.ac.uk) or, if it is more convenient, our USA Office (np-usaoffice@lancaster.ac.uk)
- For submission instructions, subscription and all the latest information visit **www.newphytologist.com**